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Transport and retention of biogenic selenium nanoparticles in biofilm-coated quartz sand porous media and consequences for elemental mercury immobilization

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17 ABSTRACT

18 Bacterial biofilms are structured cell communities embedded in a matrix of
19 extracellular polymeric substances (EPS) and a ubiquitous growth form of
20 bacteria in the environment. A wide range of interactions between biofilms and
21 nanoparticles have been reported. In the present study, the influence of a mixed
22 bacterial biofilm on retention of biogenic selenium nanoparticles (BioSeNPs) and
23 consequences for immobilization of elemental mercury (Hg^0) in a porous quartz
24 sand system were examined. BioSeNPs were significantly retained in the
25 presence of a biofilm through electrical double layer effects, hydrogen bonding,
26 and hydrophobic, steric and bridging interactions. Moreover, enhanced surface
27 roughness, pore clogging, sieving and entrapment effects mediated by the biofilm
28 also contributed to deposition of BioSeNPs. ~~Whereas~~ ^Tthiol groups associated
29 with the biofilm ^{were also involved in} ~~is a little helpful for~~ the capture of Hg^0 . It is proposed that
30 ~~oxidative complexation between~~ Hg^0 and thiol compounds or S containing
31 organic matter in the biofilm may result in the formation of Hg^{2+} -thiolate
32 complexes and HgS during the binding of Hg^0 with BioSeNPs. The formation of
33 mercury selenide was also involved in Hg^0 immobilization in the porous quartz
34 sand system.

35 **Keywords:** Biofilm, Selenium nanoparticles, Hg^0 immobilization, Nanoparticle
36 retention, Thiols, Mercury selenide

37

Introduction

A bacterial biofilm refers to the multilayer coatings of the cells surrounded by a matrix of extracellular polymeric substances (EPS), such as proteins, extracellular DNA and exopolysaccharide, attached to a living or inert surface (Smirnova et al., 2010). Such biofilms widely occur in natural environments, soil and water-sediment interfaces, and in engineered environments e.g. water treatment systems (Hullebusch et al., 2003). The prevalent formation of a biofilm is highly significant for bacterial growth and survival, and can protect them from harsh environment conditions and antimicrobial agents (Costerton et al., 1995).

Increasing commercial exploitation of nanotechnology, including applications in environmental remediation, has inevitably resulted in the release of nanoparticles (NPs) into the environment which has become an emerging environmental concern. A variety of interactions between NPs and microbial biofilms can occur but efforts to understand them have only just begun. For example, biofilms can influence the formation, adsorption, accumulation, aggregation, biodeterioration and dissolution of NPs. Furthermore, antimicrobial activities of certain NPs have been demonstrated and microbial biofilm communities can be affected (Strathmann et al., 2007; Ikuma et al., 2015). Among the possible interactions, the retention and transport of NPs in the presence of microbial biofilms has received some attention. Transport of polyvinyl alcohol-capped silver nanoparticles (AgNPs) in biofilm-coated porous media was

60 inhibited because of aggregation, ^{while} the retention of poly(vinylpyrrolidone)-
61 stabilized AgNPs in biofilm-coated sand grains was decreased (Choi et al., 2010;
62 Mitzel and Tufenkji, 2014). Tripathi et al. (2011) found that sulfate-
63 functionalized polystyrene latex particles showed increased retention in the
64 presence of a biofilm. The deposition of zero-valent iron NPs, fullerene and ZnO
65 were also all affected strongly by interactions with biofilms in porous media
66 (Tong et al., 2010; Lerner et al., 2012; Jiang et al., 2013). According to previous
67 studies, interactions between NPs and biofilms were closely associated with
68 particle surface chemistry, and the physico-chemical properties of the solution
69 (Shellenberger and Logan, 2002; Leon-Morales et al., 2007). Interactions of NPs
70 with biofilms should therefore be considered as this may not only affect the
71 environmental fate of NPs but also how nanotechnology can be effectively used
72 in environmental applications.

73

74 Biogenic selenium nanoparticles (BioSeNPs) have been demonstrated to be
75 effective for elemental mercury (Hg^0) immobilization in groundwater and soil
76 based on the reaction: $\text{Hg}^0 + \text{Se}^0 \rightarrow \text{HgSe} \downarrow$, and the effects of several
77 environmental factors such as dissolved and particulate organic matter, clay
78 minerals and EPS on Hg^0 immobilization efficiency using BioSeNPs have been
79 described in detail (Wang et al., 2018a, b; 2019a, b). BioSeNPs have been
80 reported to inhibit bacterial growth and the formation of a biofilm because of their

81 antimicrobial properties (Tran et al., 2009; Zonaro et al., 2015). However, the
82 transport and fate of BioSeNPs in the presence of microbial biofilms has not been
83 reported although this is of great interest. If Hg⁰ immobilization using BioSeNPs
84 is affected by biofilm interactions such as altered transport of BioSeNPs and Hg
85 binding, this needs to be taken into account in assessing the potential use of NPs
86 for environmental applications. In this study, the effect of a biofilm, comprising
87 a number of bacterial species isolated from natural pond water, on the transport
88 of BioSeNPs was investigated through column transport experiments as well as
89 changes in Hg⁰ immobilization efficiency.

90

2. Materials and methods

2.1 Preparation of selenium nanoparticles, groundwater and clean quartz sand

Citrobacter freundii Y9 was used to reduce sodium selenite (1 mM) resulting in the formation of biogenic selenium nanoparticles (BioSeNPs): the BioSeNPs were harvested, purified and freeze-dried according to previous publications (Cui et al., 2016; Wang et al., 2018b). The present study used filtered groundwater taken from Urumqi, Xinjiang, China: details of the filtration method and characterization of the physico-chemical properties of the groundwater are described by Wang et al. (2018b). Quartz sand, purchased from Fuchen Chemical (Tianjin, China), was sieved, washed with concentrated HCl and Milli-Q water, and oven-dried at 550 °C prior to use in experiments (Wang et al., 2018b).

2.2 Column transport experiments

Transport experiments were carried out using quartz sand porous media in the absence or presence of a bacterial biofilm. To obtain the biofilm coating, quartz sand, sterilized in a vertical heating pressure stream (LDZX-75KBS, Shanghai, China), was mixed with bacterial suspensions in a flask (at a cell concentration of $\sim 10^8$ cells ml⁻¹) and wet packed into glass cylindrical columns (10 mm inner diameter, 20 cm height) (Long et al., 2009). Bacterial suspensions were obtained by inoculation of pond water, taken near pig farms located in Urumqi, Xinjiang, China, into nutrient broth (Aoboxing Bio-tech Co. Ltd.,

Beijing, China). Wet clean quartz sand or quartz sand premixed with bacteria was added to the columns in small increments with 100 μm aperture size nylon mesh placed at the bottom of the columns. During packing, columns were vibrated gently to avoid air entrapment or layering (Tufenkji and Elimelech, 2004). For biofilm-coated quartz sand, nutrient broth was pumped through the columns (1 ml min⁻¹ flow rate) for 80 h at room temperature, and after the wet packed columns had settled for 5 h, the direction of flow was switched between upflow and downflow every 12 h to make sure that there was a uniform coating of biofilm on the sand (Liu and Li, 2008). For columns packed with clean quartz sand, a similar operation was performed using groundwater instead of nutrient broth.

After formation of the biofilm, the biofilm-coated quartz sand was subjected to community genomic analysis based on Illumina sequencing (2 × 300 MiSeq platform), the V3–4 region of the bacterial 16S rRNA genes being amplified using 341F-805R primers (Sangon Biotechnology, Shanghai, China). Clean quartz sand and biofilm-coated quartz sand were examined using scanning electron microscopy with energy-dispersive X-ray spectrometry (SEM-EDS) (Zeiss Super 55VP, Oberkochen, Germany). The zeta potential of BioSeNPs, clean quartz sand and quartz sand coated with biofilm in groundwater as a function of pH (3–11) was measured as previously described (Wang et al., 2018b).

After packing, the porosity of the columns was around 0.38. To investigate effects of the biofilm on the transport of BioSeNPs in quartz sand porous media, the columns were first pre-equilibrated with groundwater (100 pore volumes (PVs)), and then equilibrated with sonicated BioSeNPs suspensions (50 mg L⁻¹, 5 PVs) and groundwater (5 PVs). The concentration of BioSeNPs in the influent (C₀) and effluent (C) was determined at regular intervals (0.5 PVs) according to the methods described previously (Wang et al., 2018b). Finally, the clean and biofilm-coated quartz sand were harvested, freeze-dried and examined using SEM-EDS.

2.3 Hg⁰ immobilization

The influence of the biofilm in quartz sand porous media on Hg⁰ immobilization using BioSeNPs was evaluated in the same columns as used for transport experiments (Fig. S1). After the columns were packed with clean or biofilm-coated quartz sand, they were equilibrated with Hg⁰ containing groundwater (10 PVs), and then filled with 2 PVs Hg⁰-containing groundwater. The Hg⁰ containing groundwater was prepared according to an established protocol (Gai et al., 2016; Wang et al., 2018b). BioSeNPs were added to the columns at a final concentration of 50 mg L⁻¹, and after being kept for one week in a fume hood, the effluent was pumped out and collected at 0.2 PVs intervals for measurement of residual concentrations of Hg⁰ and methylmercury (MeHg).

153 Detailed information regarding Hg^0 detection using a mercury analyzer (Lumex
154 RA915+, Saint Petersburg, Russia) has been described previously (Wang et al.,
155 2018b). Hg species in solution were determined using Liquid Chromatography-
156 Hydride Generation Atomic Fluorescence Spectrometry (LC-HGAFS) (Jitian,
157 Beijing, China). After the experiments, the quartz sand with or without attached
158 biofilm was freeze-dried for SEM-EDS and ~~X-ray diffractometer~~^{ion} (XRD) (Bruker
159 D8, Karlsruhe, Germany) analysis. The biofilm-coated quartz sand with addition
160 of BioSeNPs was also characterized using X-ray photoelectron spectroscopy
161 (XPS) (Thermo Fisher Scientific, Waltham, MA, USA) ~~after experiment~~
162 according to ~~the method described in~~^{experimental} Wang et al. (2017). Another group was
163 carried out without addition of BioSeNPs ~~in clean quartz sand column~~^{to the} and the
164 quartz sand was ~~examined using XRD after experiment~~^{subsequently}. Two-way ANOVA was
165 used to examine differences in residual Hg^0 content between treatments and depth
166 of column. Differences and correlations were considered statistically significant
167 if $P < 0.05$.

168

3. Results

3.1 BioSeNPs, clean quartz sand, biofilm-coated quartz sand and groundwater

It was demonstrated previously that the BioSeNPs synthesised by *Citrobacter freundii* Y9 through bioreduction of sodium selenite were amorphous selenium nanoparticles which could be applied for effective Hg^0 immobilization (Wang et al., 2017; 2018b). The clean quartz sand particles had smooth clean surfaces (Fig. 1a), while, the surfaces of biofilm-coated quartz sand particles had a coarser appearance since the biofilm was unevenly distributed on the sand particle surfaces (Fig. 1b, c, d).

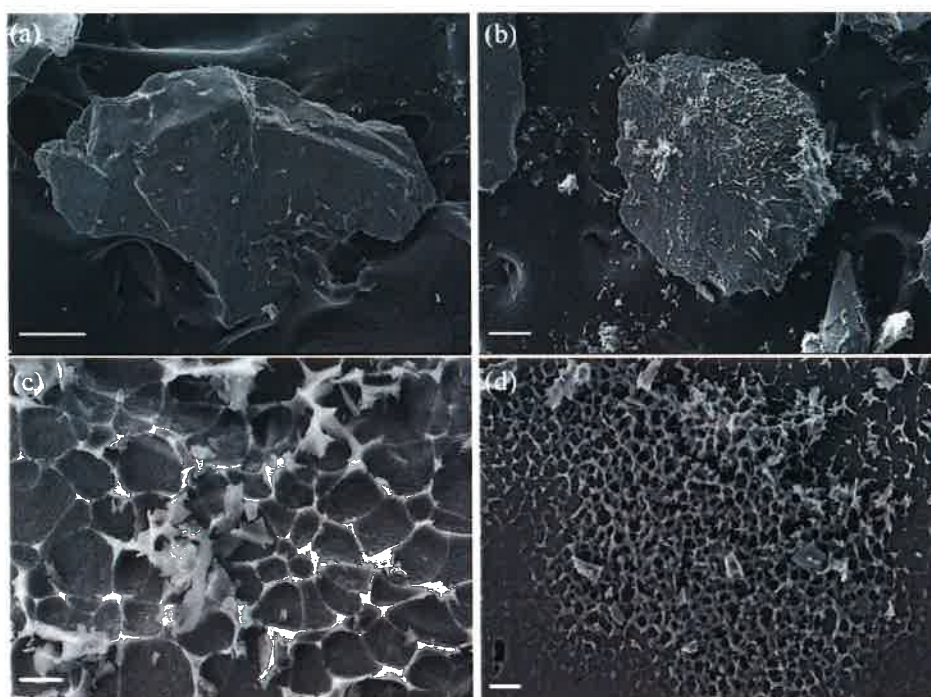


Fig. 1. SEM micrograph of clean quartz sand (a) and biofilm-coated quartz sand (b, c, d). Scale bars: (a, b) 100 μm ; (c, d) 10 μm . Typical images are shown from one of several examinations.

In this work, the biofilm was formed by a diversity of bacteria that originated from natural pond water. The taxonomic profile of the biofilm showed that the *Sphingomonas* genus was the most abundant, accounting for 14% of the total (Fig. S2). BioSeNPs, clean quartz sand and biofilm-coated quartz sand all carried a negative charge over the pH range 4-11, and the biofilm coating resulted in the quartz sand becoming less negatively charged. The magnitude of the zeta potential was in the order BioSeNPs < biofilm-coated quartz sand < clean quartz sand (Fig. S3). The physico-chemical properties of the groundwater are listed in Table S1.

3.2 Column transport experiments

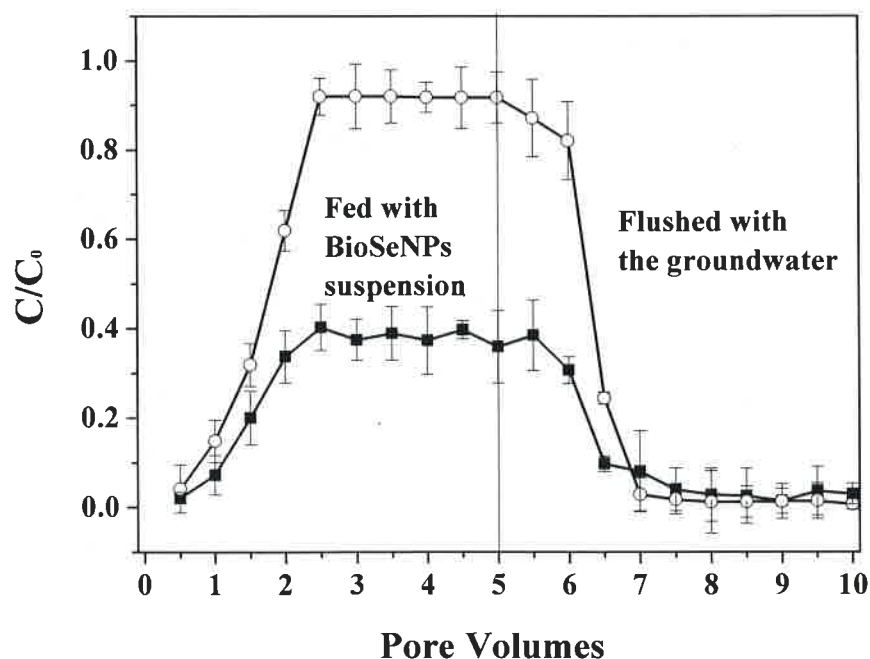
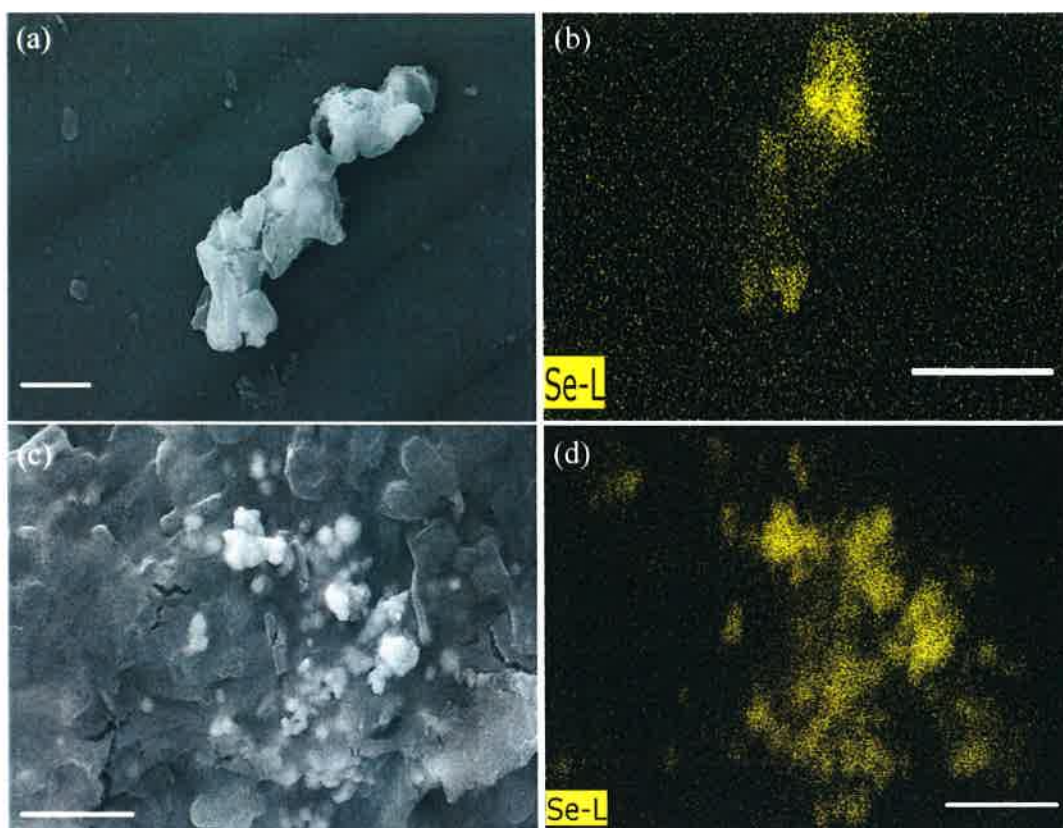


Fig. 2. Breakthrough curves of BioSeNPs in clean quartz sand and biofilm-coated quartz sand. Symbols represent: (■) biofilm-coated quartz sand; (○) clean quartz sand. Error bars represent the standard deviation (n=3).

The breakthrough curves for quartz sand with or without a biofilm coating are shown in Fig. 2. After injection with BioSeNPs suspensions, the value of C/C_0 for clean quartz sand increased gradually to ~0.91-0.92 and reached a plateau after equilibration with 2.5 PVs BioSeNPs. However, the biofilm coating lowered the steady-state breakthrough plateau to ~0.37-0.40. The amount of BioSeNPs retained in the columns was calculated through integration of the breakthrough curves. It was found that the retention of BioSeNPs in biofilm-coated quartz sand was ~1.9 times higher than for the clean quartz sand during the period of BioSeNPs injection. When flushed with groundwater, the amount of BioSeNPs leached out of the columns in the absence of a biofilm was ~1.5 times higher than the value recorded in the presence of a biofilm. In other words, BioSeNPs retention in biofilm-coated quartz sand was ~3.0 times higher than that for clean quartz sand. Less BioSeNPs remained on the smooth surfaces of clean quartz sand compared with coarser biofilm-coated sand (Fig. 3).



212

213 Fig. 3. SEM micrographs of BioSeNPs retained on clean quartz sand (a) and
 214 BioSeNPs remaining on biofilm-coated quartz sand (c). Elemental mapping of Se
 215 in clean quartz sand (b) and biofilm-coated quartz sand (d). Scale bars: 2 μm .
 216 Typical images are shown from one of several determinations.

217

218 3.3 Hg^0 immobilization

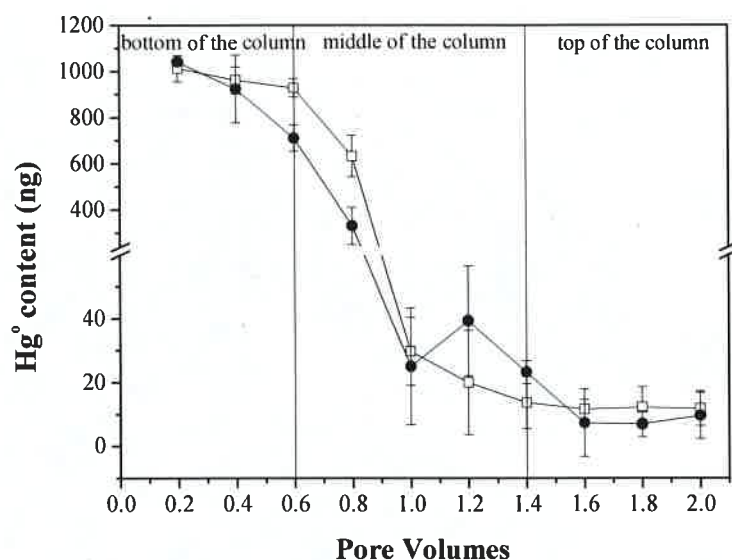


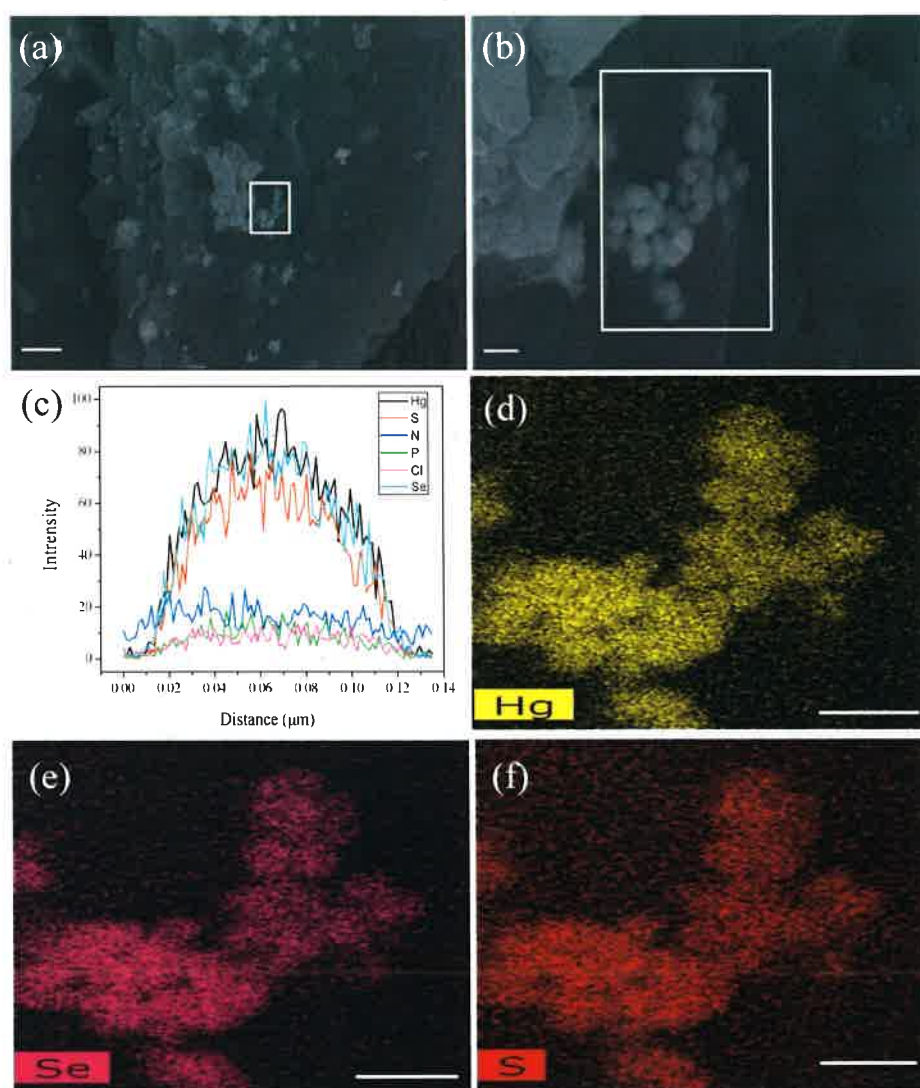
Fig. 4. Residual mercury content remaining in the quartz sand columns after treatment with BioSeNPs. Samples were taken at 0.2 PVs intervals for measurement. Symbols represent: (\square) column packed with clean quartz sand, (\bullet) column filled with biofilm-coated quartz sand. Error bars represent the standard deviation ($n=3$).

The residual Hg^0 content remaining in the groundwater after treatment with BioSeNPs in the absence or presence of a biofilm is shown in Fig. 4. ANOVA analysis revealed that the residual Hg^0 content in groundwater was significantly influenced by the biofilm and column depth ($P < 0.05$). At the bottom of the column (0.2-0.6 PVs), the residual Hg^0 content ranged from 710.7 to 1042.5 ng. However, the residual Hg^0 content significantly decreased to 13.7-632.9 ng in the middle of the column (0.8-1.4 PVs). At the top of the column (1.6-2.0 PVs), only

233 6.9-12.2 ng of Hg^0 remained in the groundwater. It was also found that the amount
234 of residual Hg^0 remaining in the column without biofilm was slightly higher than
235 that for the biofilm-coated column which means the Hg^0 capture efficiency was
236 better in the presence of biofilm than in clean quartz sand at the same depth.

237 The precipitate formed in the absence or presence of a biofilm was mainly
238 composed of Hg and Se (Fig. S4, 5) and ^{analysis} XRD ~~data~~ demonstrated the formation
239 of HgSe and HgS in the presence of biofilm (Fig. 6a). However, only HgSe was
240 ^a detected in the absence of biofilm ^a after ^{being} treated with BioSeNPs. For the biofilm-
241 coated column without addition of BioSeNPs, a HgS precipitate was also detected
242 (Fig. 6b). Small Hg^0 droplets were observed in the clean quartz sand filled column,
243 but these were not found in the biofilm-coated column. Moreover, S was an
244 obvious element occurring in the precipitate formed in the presence of the biofilm,
245 and the elemental distribution of Hg, Se and S were almost the same. Fig. 7 shows
246 the high-resolution XPS spectrum. The high-resolution spectrum of C 1s exhibits
247 three main peaks, and these binding energy peaks ^{are} ~~were~~ attributed to C-C, C-O
248 and C=O ~~/~~ respectively (Xia et al., 2019). The binding energy peak at 163.8 eV
249 ^{is} ~~was~~ assigned to RSH groups such as thiols and thiophene and ^{the} S 2p Peak at 164.9
250 eV was identified as C-S bonds (Zhang et al., 2013; Wu et al., 2017). The binding
251 energy values for Se^0 at Se 3d_{5/2} (54.5 eV) and Se 3d_{3/2} (55.2 eV) ^{are} ~~were~~ in
252 accordance with ^{values} ~~those~~ reported in the literature (Miyake et al., 1984). The
253 ^s mercury peak ^{are} at 100.3 and 104.4 eV were attributed to Hg 4f_{7/2} and Hg 4f_{5/2},

254 which demonstrating ^{es} that mercury ~~was~~ existed as oxidized mercury species
255 instead of Hg^0 (Li et al., 2018).



257 Fig. 5. Characterization of the precipitate formed in a column filled with biofilm-
258 coated quartz sand after treatment with BioSeNPs. (a, b) SEM micrograph; (c)
259 elemental line scans; elemental mapping scans of (d) Hg, (e) Se and (f) S. Scale
260 bars: (a) 1 μm; (b, d, e, f) 200 nm. Typical images are shown from one of several
261 determinations.

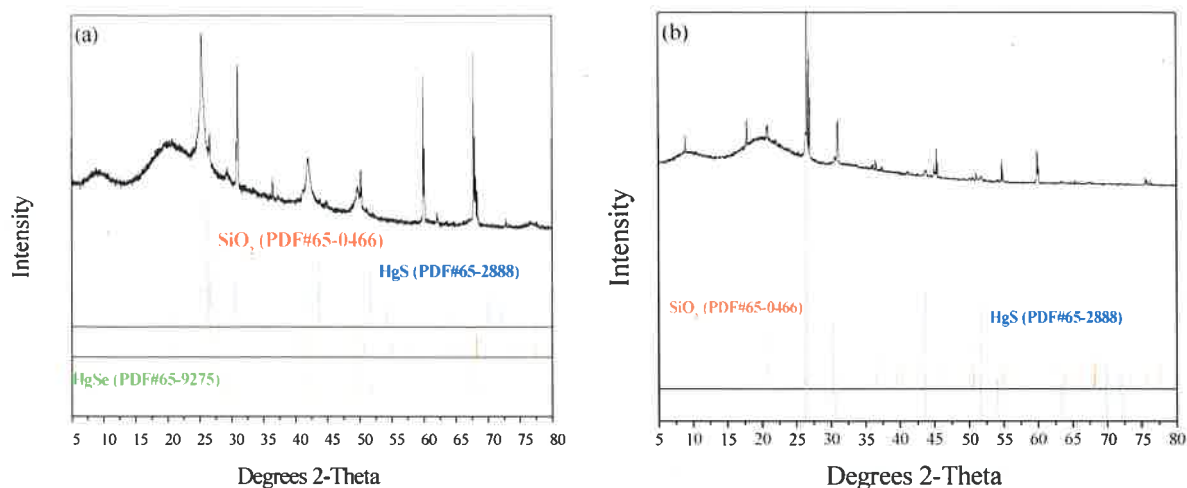


Fig. 6. XRD pattern of biofilm-coated quartz sand with (a) or without (b) addition of BioSeNPs. Typical patterns are shown.

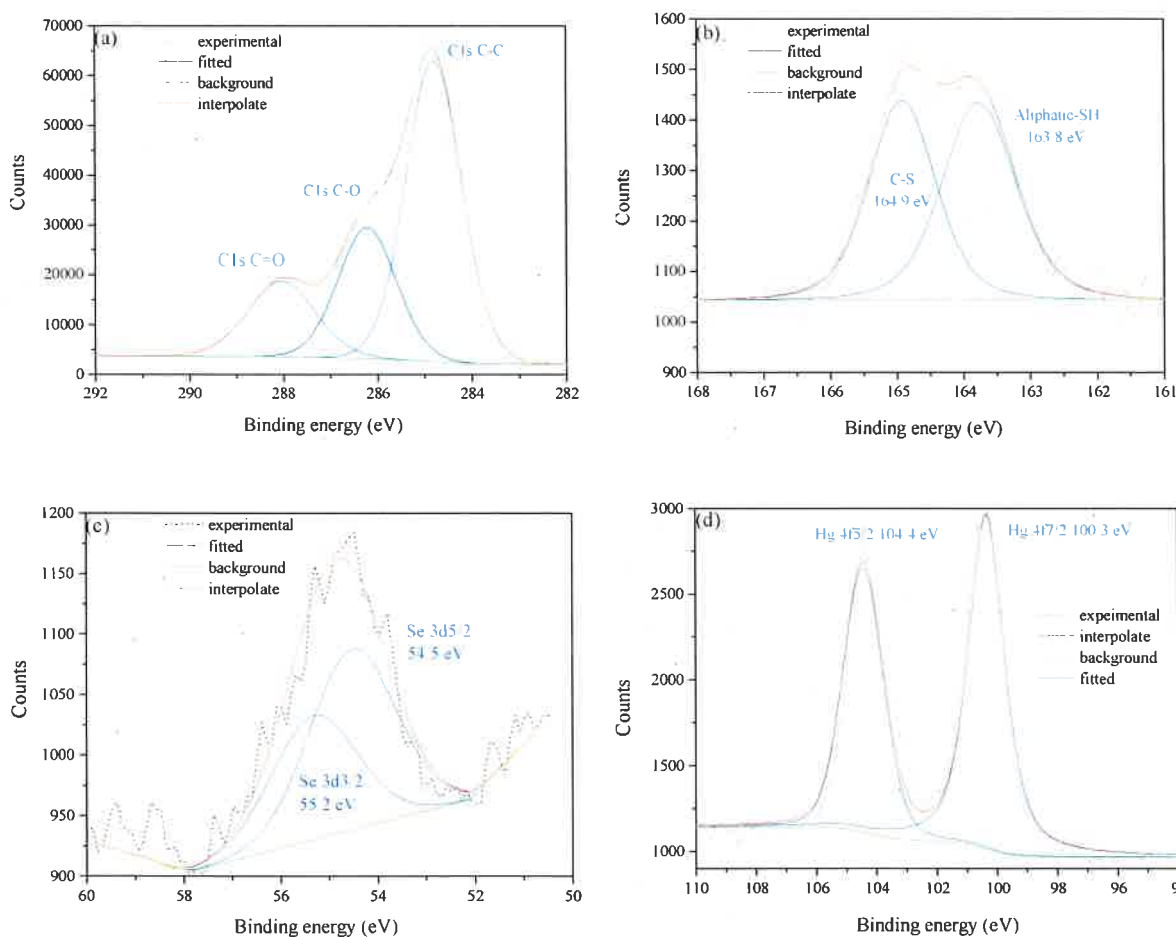


Fig. 7. High resolution XPS spectrum of the biofilm-coated quartz sand in the presence of BioSeNPs and Hg^0 . (a) C 1s; (b) S 2p; (c) Se 3d; (d) Hg 4f.

269 4. Discussion

270 Single species bacterial biofilms generally do not exist in natural
271 environments. Despite this, much previous work on biofilm effects on transport
272 of NPs in porous media has been carried out with pure culture biofilms
273 comprising one species of bacterium. In this research, the biofilm was
274 multispecies and made up of proteobacteria, bacteroidetes, acidobacteria,
275 actinobacteria, as well as other genera. The retention times for BioSeNPs in the
276 biofilm-coated column was higher than that in the absence of a biofilm, and the
277 presence of the biofilm inhibited the subsequent release of deposited BioSeNPs.
278 This is similar to previous observations where the stability of NPs such as latex
279 particles, CdSe/ZnS quantum dots, fullerene nanoparticles, zero-valent iron
280 nanoparticles and clays was reduced in the presence of a biofilm (Leon-Morales
281 et al., 2007; Tong et al., 2010; Tripathi et al., 2011; Lerner et al., 2012). Electrical
282 double layer effects, hydrogen bonding, and hydrophobic, steric and bridging
283 interactions may all affect the affinity of NPs to a biofilm as well as several other
284 physical and chemical factors (Tong et al., 2010; Lerner et al., 2012).

285

286 The porous sand media colonized by the bacterial biofilm possessed a lower
287 negative surface charge compared to control porous media without a biofilm. This
288 is in accordance with a previous report where organic coatings such as a biofilm
289 and ~~the~~ associated EPS decreased the negative surface charge of porous media

(Tripathi et al., 2011). It is known that a more negatively charged surface favours the electrical double layer which means that the negatively charged BioSeNPs encounter lower repulsion from the electrical double layer in the biofilm-coated quartz sand. In contrast, decreased retention of poly(vinylpyrrolidone)-stabilized silver NPs in a biofilm-coated sand was found to be a consequence of repulsive electrosteric forces between the biofilm EPS and poly(vinylpyrrolidone) (Mitzel and Tufenkji, 2014).

Tong et al. (2010) demonstrated that non-DLVO (Derjaguin–Landau–Verwey–Overbeek) forces played a vital role in retention of NPs in the presence of a biofilm. This was associated with ionic strength, and DLVO forces and steric interactions dominating at low and high ionic strengths, respectively (Rijnaarts et al., 1999). A considerable amount of research has reported that steric interactions can play a prominent role in controlling interactions between two polymer covered surfaces (Chen et al., 2008; Pelley et al., 2008). Such steric interactions (polymer-mediated) can be repulsive or attractive (polymer bridging), and these are dependent on the properties of the polymers and the solid surface (Jucker et al., 1998; Kim et al., 2010). Attractive polymer-mediated forces may arise through several mechanisms such as van der Waals interactions, strong ionic or biospecific ligand-receptor bonding and hydrogen bonding (Israelachvili, 2011). In the present study, BioSeNPs were coated with EPS, and the steric interaction

could be in the attractive form at certain ionic strengths because of the affinity of biofilm EPS to the polymers on the surface of BioSeNPs (Lerner et al., 2012).

Hydrophobic interactions are associated with the attachment of NPs to a biofilm, and this kind of interaction seems to be effective over greater distances (Israelachvili, 2011). It was found that bovine serum albumin-coated glass beads were more likely to retain NPs compared with alginate-coated glass beads, which are less hydrophobic, and the protein/polysaccharide ratio, an indicator of hydrophobicity and hydrophilicity, was in accordance with the retention of NPs on the biofilm surface (Xiao and Wiesner, 2013). Hydrogen bonds have also been demonstrated to be an important factor which contributes to retention of NPs. For example, hydrogen bonds mediated the interaction between polysaccharide and TiO_2 and Al_2O_3 particles (Jucker et al., 1997). It was found that *P. aeruginosa* biofilms inhibited the transport of laponite clay particles in a CaCl_2 solution, whereas transport was enhanced in a NaCl solution (Leon-Morales et al., 2004, 2007). This was because cation bridging between the surface of the biofilm and the NPs could result in more NPs deposition in the presence of Ca^{2+} (Leon-Morales et al., 2007).

EPS has been reported to account for ~85% of a biofilm mass and this greatly contributes to biofilm formation and development, and also governs attachment of NPs (Smirnova et al., 2010). It is considered that interactions between NPs and

EPS is a critical step for deposition of NPs in biofilm-coated porous media. Tong et al. (2010) showed that biofilm EPS had a great influence on C₆₀ deposition using a quartz crystal microbalance technique and dissipation experiments. The components of EPS, such as proteins, would interact with NPs through different attractive interactions (non-DLVO interactions) contributing to retention of NPs (Deguchi et al., 2007). In our previous work, EPS was found to have a great affinity for SeNPs, which would also be helpful for retention of NPs (Wang et al., 2018a).

Surface modification by the biofilm coating has also been associated with enhanced retention of NPs. The “roughness” of quartz sand was increased due to the biofilm coating and a much rougher surface would result in greater retention of NPs (Shellenberger and Logan, 2002; Liu and Li, 2008; Tong et al., 2010). One explanation is that the coating of biofilm supplied more sites of charge heterogeneity leading to increased attachment (Lerner et al., 2012). Furthermore, NPs are likely to be trapped in pores of the biofilm, pore clogging and resultant physical effects significantly contributing to retention of NPs (Ivanov et al., 2004). The formation of biofilm decreased the pore spaces and provided more pore throats and grain-to-grain contact which caused the retention of NPs, and the retention of NPs further decreased the pore space and, as a consequence, more NPs were expected to be deposited (Jiang et al., 2013).

353 It was interesting to find that the presence of a biofilm ^{was} ~~is a little~~ helpful for
 354 the capture of Hg^0 , and elemental mapping of Hg, Se and S showed almost the
 355 same distribution in the precipitate. XRD showed Hg^0 was immobilized by
 356 ^{the} BioSeNPs with formation of HgSe, which has been reported in ^{our} previous
 357 ^(Wang et al. 2017, 2018a) publications. HgS was produced according to XRD, ^{data} and ^{the} biofilm ^{ed} govern this
 358 process. XPS ^{revealed} ~~data demonstrated~~ the formation of Hg^{2+} -thiolate complexes, which
 359 ^{were} ~~was~~ associated with the facilitation of Hg^0 removal ^{the} ~~caused by~~ biofilm. In previous
 360 ^{research} ~~publications~~, ^{huge} attention has been paid to the complexation between DOM and
 361 Hg^{2+} since S-containing organic complexes have ^a strong affinity ^{for} to Hg^{2+} ~~through~~
 362 ~~chelating binding~~ (Wu et al., 2017). Organic matter rather than sulfide is more
 363 likely to complex with Hg^{2+} in anoxic environments (Ravichandran, 2004).
 364 Moreover, ^{the} ~~the~~ reactions between Hg^0 and DOM (such as thiol compounds) with
 365 ^{the} ~~formation of Hg-DOM complex~~ ^{es} ~~through oxidative complexation as expressed in~~
 366 ~~the following reactions~~ have been reported (Gu et al., 2011):
 367
$$2\text{R-SH} + \text{Hg}^0 \rightarrow \text{R-SH} \cdots \text{Hg}^0 \cdots \text{HS-R} \text{ (physicochemical sorption)}$$

 368
$$\text{R-SH} \cdots \text{Hg}^0 \cdots \text{HS-R} \rightarrow \text{R-S-Hg}^{2+}\text{-S-R} + 2\text{H}^+ + 2\text{e}^- \text{ (ligand-induced oxidative}$$

 369
$$\text{complexation)}$$

 370
$$\text{R-S-S-R}' + \text{Hg}^0 \rightarrow \text{Hg}^{2+} + \text{R-S}^- + \text{R}'\text{-S}^-$$

 371
$$\text{Hg}^{2+} + \text{R-S}^- + \text{R}'\text{-S}^- \rightarrow \text{R-S-Hg}^{2+}\text{-S-R}'$$

372 Colombo et al., (2013, 2014) found ^{the} reactivity of thiols towards Hg^0 leading ~~to~~
 373 ~~to~~ ^{to} thiol-mediated microbial oxidation of Hg^0 . Zheng et al. (2018) also showed that
 374 Hg^0 could be oxidised ^z to Hg^{2+} by natural humic acid or thiol compounds. Hg^{2+}
 375 has a high affinity with thiol-containing organic ligands such as glutathione,
 376 cysteine and dissolved organic matter, resulting in the formation of Hg^{2+} -thiolate
 377 complexes or cell-associated HgS . There are two forms of ⁱ inorganic Hg in DOM:
 378 HgS and $\text{Hg}(\text{SR})_2$. Hg^{2+} -thiolate complexes can be converted to HgS through
 379 dealkylation reactions (Manceau et al., 2015; Enescu et al., 2016; Thomas et al.,
 380 2018; Bourdineaud et al., 2019). This is in accordance with the present work ^{and} HgS
 381 ^{may be} ~~was~~ formed because of ^{such} ~~the~~ dealkylation.
 382 It was found that Hg complexed to thiolate ~~could~~ ^d reduce the bioavailability of
 383 Hg to zooplankton and fish (Sjoblom et al., 2000; Richards et al., 2011). ~~The~~
 384 ^C complexation between Hg and DOM and ^{the} formation of HgS were ^{also} found to reduce
 385 available Hg for methylation, ^{although} ~~however~~ low-molecular-weight thiols ^{can} ~~could~~
 386 mediate methylmercury (MeHg) formation by Hg methylation (Hintelmann et al.,
 387 2000; Leclerc et al., 2015; Bouchet et al., 2018). ^{However,} Whereas, MeHg and ethyl
 388 mercury (EtHg) were not detected in this work. ^{Accordingly,} ~~Accordingly,~~ ^{have} Hu et al. (2013) ^{also}
 389 reported oxidation of Hg^0 without methylation.

391 5. Conclusions

392 The presence of a biofilm enhanced the retention of BioSeNPs in quartz sand
393 porous media. The attractive interactions between ^{the}biofilm and BioSeNPs and
394 surface modification such as altered surface texture and roughness, clogging of
395 pores and sieving effects caused by ^{the}biofilm coating contributed to retention of
396 BioSeNPs. The presence of a biofilm ~~was beneficial to~~ ^{for} Hg⁰ capture due to the
397 formation of Hg²⁺-thiolate compounds, ~~through oxidative complexation and~~
398 ~~thiolate ligand~~ and HgS during the immobilization of Hg⁰ by BioSeNPs.

399

400 **Conflict of interest**

401 The authors declare no competing financial interest.

402

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Wang et al: "Transport and retention of biogenic selenium nanoparticles in biofilm-coated quartz sand porous media and consequences for elemental mercury immobilization"

Thank you for the reviewers' comments on our manuscript: they are of great help for revising and improving our paper. We have studied the comments carefully and have made the relevant corrections. All revisions are marked in red in the revised manuscript. Detailed responses to the reviewers' comments are as follows (the replies are highlighted in blue):

Reviewer #1:

The influence of a mixed bacterial biofilm on retention of BioSeNPs was investigated in this paper. Meanwhile, the effect of it was found that the biogenic selenium nanoparticles in biofilm-coated quartz sand porous media could immobilize the elemental mercury from water. Furthermore, thiol groups could enhance the mercury immobilization. However, it is well known that mercury could react with Se or thiol. So, what are the new phenomenon, mechanism or technology in this paper?

Response: It is known that mercury could react with Se or thiol, however, the influence of a mixed bacterial biofilm on retention of biogenic selenium nanoparticles (BioSeNPs) and consequences for immobilization of elemental mercury (Hg^0) in a porous quartz sand system has not been reported. In the present study, the transport of

BioSeNPs was inhibited by biofilm, but the oxidative complexation between Hg^0 and thiol and the formation of HgS is a little helpful for the capture of Hg^0 .

1. In Figure 7, through the comparison between the clean quartz sand and biofilm coated quartz sand, the efficiencies of Hg removal is not obvious, and it is said that thiol groups associated with the surface of the biofilm, the author also pointed that oxidation of Hg^0 to Hg^{2+} by thiol compounds, so Se NPs and thiol groups don't have very good performance for the removal of mercury by this biofilm.

Response: BioSeNPs have a very good performance for the immobilization of Hg^0 . ~98.8% of the Hg^0 was captured in the top of the column. The Hg^0 immobilization using BioSeNPs in the presence/absence of biofilm in the bottom of the column were both inhibited, only 2.9-31.9% of the Hg^0 were captured since the transport of BioSeNPs was suppressed as shown in breakthrough curves. However, at the bottom of the column, the oxidative complexation between Hg^0 and thiol compounds in biofilm and the formation of HgS is a little helpful for the capture of Hg^0 , compared with clean quartz sand filled column.

2. This work should give the XPS spectrum of the BioSeNPs and after adsorption with Hg^0 . The S2p XPS is also needed.

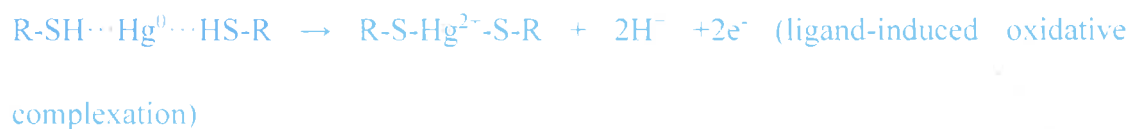
Response: We have characterized the BioSeNPs using SEM-EDS, XRD and TEM in our previous, the particles were amorphous elemental selenium nanoparticles. XRD and XPS were used to demonstrate the formation of HgSe after adsorption of BioSeNPs with Hg^0 . In this study, we added XPS spectrum (Fig. 7) according to reviewer's opinion. S 2p high resolution XPS spectrum was also supplied to reveal the

occurrence of thiolate compounds. XRD pattern (Fig. 6) was also given to demonstrate the formation of HgSe and HgS.

3. The Hg⁰ adsorption isotherm should be given.

Response: The immobilization of Hg⁰ using BioSeNPs has been illustrated in our previous work and introduction based on the reaction: $\text{Hg}^0 + \text{Se}^0 \rightarrow \text{HgSe} \downarrow$

The mechanisms about oxidative complexation between Hg⁰ and thiol were given as follows:



4. The whole article is not easily understood.

Response: We corrected the whole manuscript carefully to make it easily understood.

Reviewer #2:

The authors report transport and retention of biogenic selenium nanoparticles in biofilm-coated quartz sand porous media and consequence for elemental mercury immobilization. The method is short of novelty. Thus, it is not acceptable.

Response: Although methods used in the present study are usual or regular methods, the effects of biofilm on BioSeNPs and consequences for Hg⁰ immobilization have not been reported. Biofilm is widely spread in the environment, it necessary to evaluate the effects of biofilm which is helpful to well understand the applications of

BioSeNPs for Hg^0 immobilization. It was found that the transport of BioSeNPs was inhibited, but the oxidative complexation between Hg^0 and thiol and the formation of HgS is a little helpful for the capture of Hg^0 .

Reviewer #3:

Because biofilm is omnipresent in soil and water environments, it is important to evaluate the effects of biofilm in retention of nanoparticles and consequences for heavy metal immobilization.

The paper "Transport and retention of biogenic selenium nanoparticles in biofilm-coated quartz sand porous media and consequences for elemental mercury immobilization" by Wang et al. is an interesting work with some novel findings, which answered the important questions in the field of bioremediation using nanoparticles. In this work, the authors demonstrated that BioSeNPs were significantly retained by biofilm and the underlying mechanisms were proposed from the perspective of their chemical interaction, physical impediment of transport of NPs. The consequences of BioSeNPs by biofilm for their sequestration capacity of elemental mercury was also evaluated. All these findings are helpful for improvement of bioremediation of Hg using Se nanoparticles.

Generally speaking, this paper was very clearly organized and well written. Its publication can attract wide attention from the scientists in bioremediation and other relevant fields.

Some speculative sentences should be removed from the section of conclusions.

Response: Many thanks for your estimation and we have revised conclusions.

Reviewer #4:

The authors have conducted study on "Transport and retention of biogenic selenium nanoparticles in biofilm-coated quartz sand porous media and consequence for elemental mercury immobilization". The article however can be made more scientifically strong by incorporating some additional analytical studies on nanoparticles characterization and impact of retention of selenium nanoparticles and mercury immobilization of biofilm microbial community.

Some of my following comments may help authors to rework on the MS for possible publication on this subject. Hence the paper needs major revision.

Specific comments

1. It is advisable to further investigate nanoparticles using TEM, XRD etc. after microbial synthesis and after they are immobilized on biofilm. SEM can't revealed information nanoparticles.

Response: In our previous papers we have investigated BioSeNPs using TEM and XRD, and the synthesized nanoparticles were demonstrated to be amorphous selenium nanoparticles which have strong affinity to Hg^0 . Thus, we did not add results of TEM and XRD in the present study.

After immobilization, we used XRD and XPS to exam the product. It was found HgS was synthesized in the presence of biofilm with/without addition of BioSeNPs according to XRD. XRD results showed HgSe and HgS were formed in the presence of biofilm after addition of BioSeNPs. XPS spectrum demonstrated that Hg-thiolate complex was also occurred in the presence of biofilm after addition of BioSeNPs.

2. Please do further characterization study to verify formation of Hg^{2+} - thiol complex and mercury- selenide complex.

Response: We added XRD (Fig. 6) and XPS (Fig. 7) analysis to verify formation of Hg^{2+} -thiol complex and mercury- selenide complex.

3. It will be highly useful if author will perform some toxicity study of synthesized selenium nanoparticles on different ecotoxicological model organisms as these nanoparticles are proposed to be used in environmental remediation. Therefore interaction of these nanoparticles with aquatic organism will greatly enhance applicability of present study.

Response: Your advice is very useful, although BioSeNPs are considered as unarmful to organisms, it is still necessary to carry out the toxicity study of BioSeNPs on different ecotoxicological model organisms. We will do this work systematically in the next paper.

4. Changes in nanoparticles properties (size, stability etc.) after they come in contact with natural water sample may be also useful aspect of study.

Response: Yes, it is of great help to take the changes of nanoparticles' properties into consideration after they were added into natural water. In our previous work, we have investigated the effects of extracellular polymeric substances, dissolved and particulate organic matter, clay minerals and salinity on the stability of BioSeNPs. We will continue to do this kind of work in future.

In conclusion, we have attended to all referee's comments and amended the manuscript accordingly. We have justified our approach and conclusions in the detailed responses. We hope the manuscript is now acceptable for publication.

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